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Effects of global change on insect pollinators: ļ multiple drivers lead to novel communities

Nicole E Rafferty^{1,2}

- 5 Global change drivers, in particular climate change, exotic
- species introduction, and habitat alteration, affect insect 6
- pollinators in numerous ways. In response, insect pollinators 7
- show shifts in range and phenology, interactions with plants 8
- and other taxa are altered, and in some cases pollination 0
- services have diminished. Recent studies show some 10
- 11 pollinators are tracking climate change by moving latitudinally
- and elevationally, while others are not. Shifts in insect pollinator 12
- phenology generally keep pace with advances in flowering, 13
- although there are exceptions. Recent data demonstrate 14
- competition between exotic and native bees, along with rapid 15
- positive effects of exotic plant removal on pollinator richness. 16
- Genetic analyses tie bee fitness to habitat quality. Across 17
- drivers, novel communities are a common outcome that 18
- deserves more study. 19

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Introduction 29

Global change is affecting insect pollinators in profound 30 ways. Climate change, exotic species introduction, and 31 habitat loss are affecting all major aspects of the biology of 32 insects that pollinate plants in both natural and agricul-33 tural communities, altering their distribution, phenology, 34 abundance, physiology, and morphology [1-5]. The con-35 sequences of these effects are complex, perturbing plant-36 pollinator interactions in subtle but important ways and in 37 some cases resulting in local extinction [2]. Despite the 38 complexity, understanding these consequences is critical: 39 just as the vast majority of flowering plants depend on 40 insects for pollination [6], we rely in large part on insects 41 to pollinate our crops, a valuable ecosystem service [7]. 42

Among the many insect taxa that serve as pollinators, 43 bees, flies, butterflies, and moths have received the most 44 study in the context of global change. Within these taxa, 45 bees are key pollinators of both crop plants and wild 46 plants [8], and studies on bees have dominated the 47 literature on plant-pollinator interactions under global 48 change. Because bees rely heavily on floral resources both 49 for their own sustenance and to provision their offspring, 50 their fitness is strongly determined not only by the direct 51 effects of global change but also by the influence of global 52 change drivers on flowering plants. 53

Here, I consider the effects of several global change 54 drivers on insect pollinators, with an emphasis on what 55 we know about the effects on native bees. First, I discuss 56 how climate change is affecting insect pollinators, as this 57 is a topic of active research that illustrates a suite of 58 responses. Second, I review the effects of exotic species, 59 both insect and plant taxa, on insect pollinators. Third, I 60 consider another global change factor, habitat alteration 61 and loss, and its effects on insect pollinators. Through-62 out, I consider both direct effects on pollinators and 63 effects that are mediated via plants and other interspe-64 cific interactions. Given biotic pollination is by definition 65 a multitrophic interaction, greater consideration of how 66 global change alters species interactions is needed to 67 improve conservation and management of pollination 68 services. 69

Effects of climate change

The responses of insect pollinators to climate change 71 have been relatively well-studied, although much 72 remains to be resolved. For the most part, experimental 73 studies of climate change factors on insect pollinators 74 have focused on temperature [9-12], an important deter-75 minant of developmental rate [13]. Manipulations of 76 other factors, such as carbon dioxide [14] or precipitation 77 [15], have been applied to plants with subsequent mea-78 sures of pollinator responses to altered floral traits. Com-79 plementing experimental approaches are long-term data, 80 historical observations, and museum specimen records 81 that can be correlated with ambient temperatures and 82 other climate variables to describe insect responses 83 [1, 16].84

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Among the most striking consequences of climate change 85 have been shifts in the spatial distributions of insect 86 pollinators. Given the rapid life cycles and high mobility 87 of most insect pollinators, are they able to keep pace with 88 anthropogenic climate change by tracking environmental 89

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conditions over space? Evidence is mixed. On the one 90 hand, Kerr et al. [4^{••}] discovered bumble bees (Bombus 91 spp.) across two continents have not tracked warming 92 temperatures, as evidenced by a failure to expand their 93 northern latitudinal range limits. On the other hand, 94 several studies have shown that bumble bees have moved 95 upward in elevation in montane ecosystems [4^{••},17,18], 96 and some butterflies have shifted up in altitude [19]. Both 97 a nymphalid butterfly (Polygonia c-album) and a lycaenid 98 butterfly (Aricia agestis) in Britain have greatly expanded 99 100 their ranges northward in association with warming [20,21]. A key question that has been not been considered 101 for most taxa is how these spatial shifts affect interactions 102 with floral resources and thereby influence both pollinator 103 fitness and patterns of pollen flow and reproductive 104 output of plants. Differential shifts among taxa will 105 almost certainly translate into modified communities, 106 especially as perennial plants are likely to lag behind 107 108 their pollinators. In addition, it remains largely unknown whether traits or phylogenetic relationships can explain 109 variable spatial responses among taxa (but see [4^{••},22]). 110 To understand constraints on the distributions of insect 111 pollinator populations and predict how distributions will 112 be affected by climate change directly and via effects on 113 host plants and other species with which pollinators 114 interact, species distribution models can be a useful tool 115 [23,24]. 116

Shifts in the phenologies of insect pollinators are another 117 conspicuous signal of climate change. Multiple species of 118 bees have significantly advanced their phenologies [1], as 119 have many butterflies and moths [25,26]. Among lepi-120 dopterans, variable responses can be partially explained 121 by traits such as diet breadth [26]. In contrast to spatial 122 shifts, the consequences of climate change-induced tem-123 poral shifts for plant-pollinator interactions have received 124 much attention. Community-level analyses indicate bees 125 and the plants they pollinate are advancing at similar rates 126 [1], whereas butterflies and their nectar sources show 127 different sensitivities to temperature [27[•]]. In general, 128 experimental studies suggest phenological mismatches 129 are unlikely to lead to complete decoupling of interac-130 tions among insect pollinators and plants [28,29]. In part 131 132 this outcome is not surprising: plant-pollinator interactions tend to be generalized [30] and nested, with 133 specialists interacting with generalists [31], and high 134 rates of interaction turnover [32]. However, there are 135 examples of specialized plant-bee interactions that are 136 likely becoming disrupted as phenologies shift [33,34]. 137 Even subtle phenological mismatches are likely to have 138 consequences for interaction strengths, fitness, and the 139 evolution of life histories [35]. Whereas the conse-140 quences of mismatches for plants have been commonly 141 measured in terms of seed production [29,36], the con-142 sequences for pollinators have gone unquantified [37]. 143 Also in contrast to the situation for insect pollinator 144 phenology, where few studies have linked responses 145

to traits or phylogenies, flowering phenology responses 146 to climate change have been associated with traits such 147 as flowering season, life history, and pollination mode 148 [38,39] and exhibit phylogenetic signal across continents 149 [40]. Together, these gaps in understanding point to a 150 need for more studies at the community level: a com-151 munity approach should simultaneously create opportu-152 nities for trait-based analyses and enable the conse-153 quences of phenological mismatches from the 154 pollinator perspective to be quantified. 155

Other aspects of climate change that have been demon-156 strated to affect insect pollinators via flowering plants 157 include elevated carbon dioxide and decreased precipi-158 tation. Plants grown under elevated carbon dioxide can 159 have altered floral traits, such as nectar composition [14] 160 and pollen protein concentration [41]. In turn, these 161 altered traits can influence the fitness of insect pollinators; 162 Hoover et al. [14] found that Bombus terrestris workers 163 exhibited reduced longevity when fed synthetic nectar 164 mimicking that of flowers produced under elevated car-165 bon dioxide, and Ziska et al. [41] posit that reduced 166 protein in goldenrod pollen could negatively affect bees. 167 Experimental drought had variable effects on floral vola-168 tiles but consistently reduced flower size and floral display 169 across four species, resulting in different communities of 170 bees, flies, and butterflies visiting the flowers in the 171 drought treatment [15]. In general, a tight link between 172 the direct effects of climate change on floral resources and 173 the consequent effects on insect pollinators has yet to be 174 made. In part, this is because it is difficult to isolate the 175 effects of complex floral responses on mobile insects, 176 particularly in the field and at the population and com-177 munity levels. As molecular genetic techniques and tech-178 nologies that allow automated identification of individual 179 bees, for example as they pass over radio frequency 180 identification readers, are refined, larger-scale field-based 181 studies of pollinator fitness and foraging responses should 182 become more feasible. 183

Effects of exotic species

Human-aided transport and introduction of exotic species 185 is a major driver of global change, reshaping fundamental 186 ecological relationships [42]. Focusing in on exotic insect 187 pollinators, we know the most about the impacts of non-188 native bees on native bees [43]. Non-native bees include 189 long-established domesticated honey bees (Apis melli-190 fera), more recently-introduced commercial pollinators, such as *Bombus terrestris* [44], and accidental introductions 191 of species such as *Hylaeus communis* [45]. Alien pollinators 192 can compete with native pollinators for resources, poten-193 tially reducing their fitness, altering patterns of pollen 194 flow, and ultimately changing community structure to the 195 disruption of ecosystem services [46,47]. Not surprisingly, 196 the best-studied interactions between exotic and native 197 bees involve honey bees. Building on prior experimental 198 work that demonstrated competition for floral resources 199

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- ²⁰⁰ between honey bees and a native bumble bee (*B. occidentalis*; [48]), Thomson [49^{••}] used a 15-year-long data
- set to show a negative relationship between feral A. 201 mellifera densities and Bombus spp. densities. Similarly, after honey bees invaded a tropical reserve, solitary bees 202 were observed to visit different plant species because of 203 competition, but declines in the native bees were not 204 detected [50]. Thus, the effects of exotic insect pollina-205 tors on native pollinators likely depend on factors that 206 modify the strength of competition, such as niche overlap 207 and flexibility, as well as interacting effects of other 208 stressors, such as drought, that modulate floral resource 209 availability. 210

Turning briefly to non-native plants, several studies 211 have investigated how exotic plants influence plant-212 pollinator interactions [46,51–53]. Recently, a large 213 experiment by Kaiser-Bunbury et al. [54**] showed 214 exotic plant removal resulted in about 20% more polli-215 nator species in restored sites, with more generalized 216 plant-pollinator networks and higher fruit set of com-217 mon species. These results suggest removal of non-218 native species can rapidly enhance pollinator richness 219 but may, as the authors note, hinge on nearby popula-220 tions of pollinators to colonize restored sites [54^{••}]. More 221 broadly, no real consensus on the effects of exotic plants 222 on insect pollinators has emerged, with both positive and 223 negative effects reported [46,51]. Moving forward, 224 greater integration of the study of exotic species with 225 226 the study of phenological and range shifts, which can 227 similarly modify interaction strengths and create novel communities, would be productive. 228

229 Effects of habitat alteration and loss

Habitat alteration and loss is widely recognized as a 230 contributor to declines of insect pollinators [55]. 231 Changes in land use are associated with changes in 232 pollinator community composition and richness; in par-233 ticular, conversion to arable land is associated with 234 declines in bee and wasp species richness over 80 years 235 in Britain [56]. Agricultural intensification carries its own 236 suite of effects on insect pollinators, including the direct 237 effects of pesticides such as neonicotinoids, which can 238 have multiple debilitating effects on bees [57–59], weak-239 ening pollination services [60]. Using genetic analyses, a 240 recent study by Carvell *et al.* [61[•]] showed that lineage 241 survival of three bumble bee species increased as a 242 function of nearby high-quality foraging habitat, quanti-243 244 fied as semi-natural vegetation, spring floral resources for queens, and overall flower cover in spring and summer. 245 Bumble bee nesting density also can be negatively 246 related to the percent of paved surface and positively 247 related to the amount of natural oak woodland-chaparral 248 habitat [62]. 249

- 250 Some traits serve as predictors of the severity of effects of
- ²⁵¹ habitat alteration and loss on insect pollinators. Generally,

specialized pollinators are more sensitive to land use 252 impacts [63,64]. Within bees, a global analysis indicated 253 stronger negative effects of overall agricultural intensifi-254 cation and isolation from natural habitat for species that 255 nest above ground, whereas species that nest below-256 ground were adversely affected by land tilling [65]. 257 The abundance of social bees was also more negatively 258 affected by isolation than was the case for solitary bees 259 [65]. Some pollinators may be able to adjust their foraging 260 distances in response to landscape-scale variables, as seen 261 with bumble bees capable of foraging farther to find 262 patches of greater floral diversity in landscapes that are 263 relatively homogeneous [62]. Altogether, multiple studies 264 indicate that ecological intensification practices, such as 265 increasing floral resource availability and diversity across 266 landscapes, have positive effects on insect pollinator 267 persistence in the face of habitat alteration [66]. Never-268 theless, with changing land use, pollinator behavior and 269 species composition are likely to change, modifying inter-270 actions and pollination services. 271

Conclusions

As we become increasingly aware that species interactions 273 shape species distributions in time and space and modu-274 late the direct effects of global change, considering insect 275 pollinators in a community context should be a priority. 276 For example, Forrest and Chisholm [67[•]] demonstrated 277 that warmer temperatures led simultaneously to higher 278 rates of activity and nest provisioning by a solitary bee 279 (Osmia iridis) and to increased rates of brood parasitism by 280 a wasp (*Sapyga* sp.). Thus, positive effects of warming are 281 likely to be negated by altered interaction frequencies 282 with a natural enemy [67[•]], a result that would not be 283 predicted in isolation of community context. Community-284 level analyses also detect broader trends before pairwise 285 interactions are disrupted or individual species decline. 286 For example, a study of phenological overlap in 287 Greenland over 18 years points to disrupted plant-polli-288 nator interactions as the flowering season shrinks, poten-289 tially leaving pollinators without floral resources late in 290 the season [68]. 291

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Much progress has been made in understanding the 292 effects of individual global change drivers on insect 293 pollinators. Moving forward, further progress in under-294 standing and mitigating anthropogenic disturbances 295 could be made by searching for common outcomes across 296 drivers. All three of the global change drivers highlighted 297 here are likely to result in novel interactions and commu-298 nities. Climate change, for example, alters overlap among 299 species via spatial and temporal shifts, among other 300 mechanisms. Introduced exotic species interact with 301 native species in novel ways. And habitat alteration 302 and loss can result in novel species composition and cause 303 species to modify behavior, altering interactions. By tying 304 these common outcomes to resulting eco-evolutionary 305 dynamics, we can begin to anticipate how global change 306

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will reshape insect pollinator communities and pollina-307 tion services. 308

Conflict of interest statement 309

Nothing declared. 310

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 This paper demonstrates the important insights to be gained by long time
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